

# Surface modification of sodium alginate-polyvinyl alcohol hydrogel beads using low-pressure cold plasma and application for methylene blue removal from water

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## ABSTRACT

This study investigates the adsorption efficiency of methylene blue (MB) from aqueous solutions using sodium alginate-polyvinyl alcohol (SA/PVA) hydrogel beads modified through low-pressure cold plasma (LPCP) treatment. The LPCP process significantly enhanced the surface properties of the hydrogel beads, improving their adsorption capacity from 27.48 % (untreated) to 53.03 % (P15 treatment) at pH 8.2. Plasma-treated beads exhibited substantially higher MB removal compared to untreated beads, with the highest adsorption capacity of 175.506 mg g<sup>-1</sup> under the P15 treatment condition. Adsorption followed the Freundlich model with high correlation coefficients ( $R^2 = 0.998$ ), suggesting multilayer adsorption on a heterogeneous surface. Adsorption kinetics adhered to a pseudo-second-order model ( $R^2 > 0.98$ ), indicating chemisorption as the dominant mechanism. Thermodynamic analysis demonstrated spontaneous and endothermic adsorption ( $\Delta G^\circ = -7.55 \text{ kJ mol}^{-1}$  at 333 K). Plasma-treated hydrogel beads showed excellent reusability, maintaining over 85 % adsorption efficiency after three cycles. These findings highlight the potential of LPCP-treated hydrogel beads as an effective and sustainable solution for MB removal in wastewater treatment applications.

## 1. Introduction

Industrial growth in almost all sectors, especially textile, and paper, where dyes are routinely added to most products, is one of the most common causes of water pollution when untreated waste flows into land resources or directly back into aquatic systems [1]. The presence of these dyes in water bodies leads to contamination, a decline in water quality, and the introduction of toxic substances, posing severe threats to aquatic organisms and human health. Addressing this environmental challenge is critical to safeguarding ecosystem integrity and public health [2]. Water pollution caused by the textile, pharmaceutical, and chemical industries is a significant environmental concern. Synthetic dyes stand out among these pollutants due to their efficiency and persistence. The global textile industry releases over 280,000 metric tons of synthetic dyes into aquatic environments annually [3]. Contamination with

methylene blue (MB) in the field, particularly in large quantities or uncontrolled environments, will have broad effects such as aquatic toxicity, soil contamination, human health risks, and water contamination [4]. MB can impair the oxygen-carrying capacity of aquatic organisms, potentially causing suffocation. Additionally, MB can alter microbial populations in soil, disrupt nitrogen fixation, and negatively impact plant growth [5]. MB also can negatively influence human health; it is toxic in people and can lead to health problems in people. It may cause skin, eye, and respiratory irritations. Higher doses of it may result in side effects like nausea, vomiting, diarrhea, and, in extreme cases, hemolytic anemia [6]. MB presents a multidimensional ecological and human health challenge of scientific significance. At the molecular level, the MB defines toxicological traits like mutagenic potential, damage to cellular membranes, and disturbance of biochemical pathways. Experimental studies in the laboratory revealed that MB can

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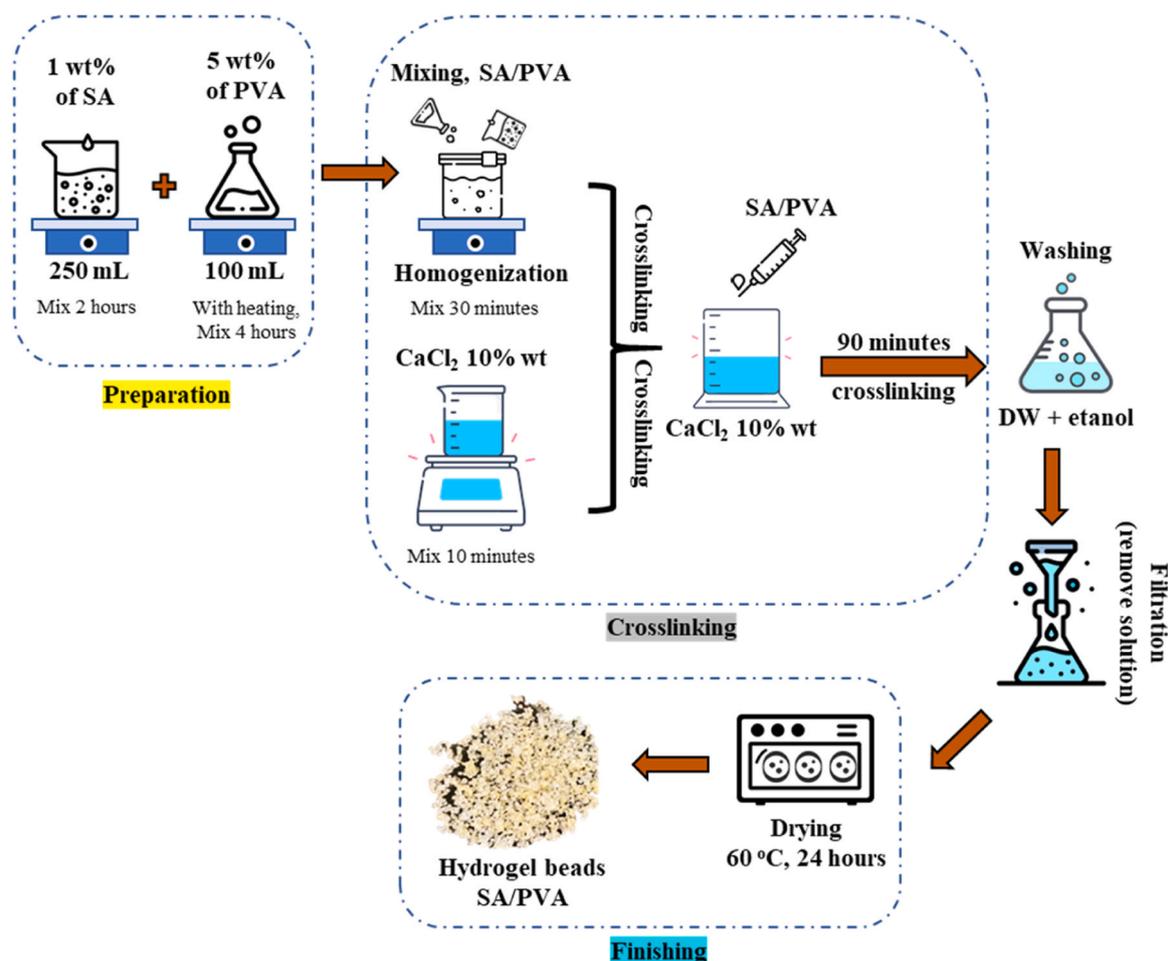


Fig. 1. Preparation of SA/PVA Hydrogel Beads.

induce oxidative stress and generate reactive oxygen species in biological systems, potentially leading to chromosomal aberrations [7].

Several approaches have been developed to reduce MB contamination, including cautious handling, proper storage in closed containers, and advanced waste management strategies. Various methods for removing MB have been examined, including adsorption, coagulation-flocculation, and membrane filtering [8]. The complex molecular structure of MB makes it quite resistant to conventional water treatment methods due to its chemical stability, strong intermolecular interactions, and persistent nature in aquatic environments [9]. Common physico-chemical treatment techniques, such as coagulation, flocculation, or adsorption, can remove only about 60–70 % of dyes, leaving substantial pollution in water systems [10].

Among these techniques, adsorption is a promising approach due to its cost-effectiveness, high efficiency, and environmental sustainability. Adsorption operates under mild conditions, requires minimal energy input, and can be applied to a wide range of adsorbents [11,12]. Several adsorbent materials have been reported for MB removal from water, including biochar [13,14], activated carbon [15,16], zeolites [17], clay minerals [18,19], metal-organic frameworks (MOFs) [20,21,22], and polymer-based adsorbents [23,24,25].

Hydrogel-based adsorbents have gained significant attention due to their high water retention, tunable porosity, and functional group availability, which enhance dye adsorption [26]. Among various hydrogel systems, sodium alginate (SA) and polyvinyl alcohol (PVA) are widely used due to their biocompatibility, non-toxicity, and ability to form stable hydrogel networks. SA, a natural polysaccharide, provides a biodegradable matrix with abundant carboxyl (-COOH) groups that facilitate MB binding. PVA enhances the hydrogels mechanical stability,

elasticity, and reusability, making it an ideal component for adsorption applications [27].

Surface modifications are often required to improve the adsorption efficiency further. A promising technique for enhancing surface properties is cold plasma (CP) treatment, which induces physical and chemical modifications without altering the bulk properties of the material [28,29]. CP is an emerging surface modification technique that utilizes a partially ionized gas composed of energetic electrons, ions, free radicals, and reactive oxygen and nitrogen species (RONS) [30]. These reactive species interact with material surfaces, inducing both physical and chemical modifications without significantly altering the bulk properties [31]. One of the primary effects of CP treatment is the introduction of oxygen-containing functional groups, such as hydroxyl (-OH), carboxyl (-COOH), and carbonyl (C=O), which enhance surface hydrophilicity and create additional active adsorption sites [32]. These modifications improve electrostatic interactions between the adsorbent and target contaminants, such as methylene blue (MB), thereby enhancing adsorption performance. Furthermore, CP can alter surface morphology by introducing nano-scale roughness, increasing the available surface area for adsorption [33].

While CP has been widely applied for modifying various materials, its operational conditions vary depending on the plasma source, pressure, and gas composition. Atmospheric-pressure cold plasma (APCP) is commonly used due to its simplicity and scalability; however, it often suffers from non-uniform plasma distribution and limited control over reactive species, leading to inconsistent surface modifications [34]. In contrast, low-pressure cold plasma (LPCP) operates under vacuum conditions (typically below 100 Pa), allowing for better plasma uniformity, controlled reactive species generation, and deeper surface

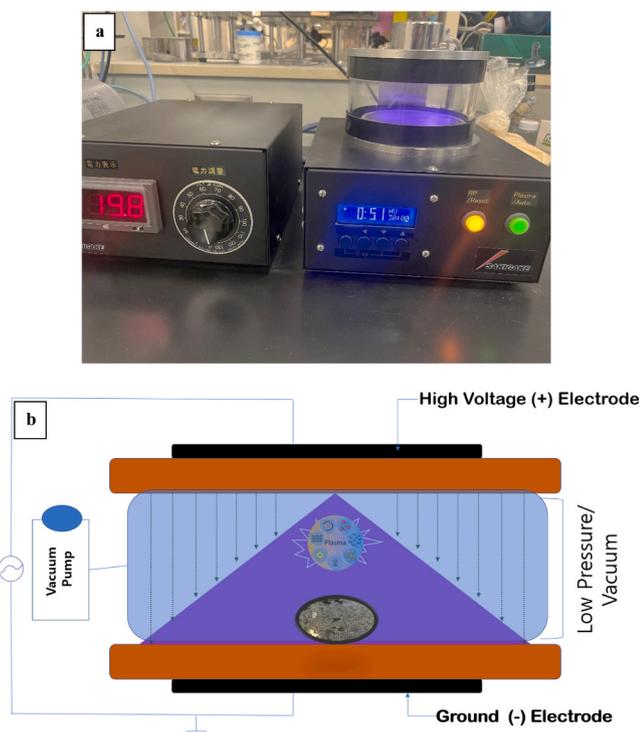


Fig. 2. (a) Vacuum plasma device (Sakage, YHS-R) and (b) schematic diagram of LPCP treatment for hydrogel beads.

penetration [35]. These advantages make LPCP particularly effective for modifying polymeric materials, including hydrogels.

Despite its potential, LPCP remains underexplored in hydrogel surface modification, particularly for water treatment applications. Most studies on CP-modified adsorbents have focused on carbonaceous materials such as activated carbon and biochar, while studies utilizing LPCP for hydrogel-based adsorbents remain limited [36]. In contrast, conventional chemical modification techniques, such as acid/base treatment, grafting polymerization, or functionalization with crosslinking agents, have been widely employed to introduce functional groups onto hydrogel surfaces [37,38]. However, these chemical methods often involve using toxic reagents, generating chemical waste, and may lead to unwanted structural degradation of the hydrogel matrix. In this study, we investigated the relationship between LPCP treatment conditions such as varying treatment times and the efficiency of MB removal from water. The findings of this study on the surface modification of SA/PVA hydrogel beads using LPCP aim to advance the development of more efficient water treatment technologies for removing MB from aqueous solutions.

## 2. Materials and method

### 2.1. Materials

Sodium alginate (SA) was purchased from Wako Chemicals, Tokyo, Japan. Calcium chloride ( $\text{CaCl}_2$ ,  $\geq 96.0\%$ ), Hydrochloric acid (HCl, 36.0–38.0%), methylene blue (MB), and sodium hydroxide (NaOH,  $\geq 96.0\%$ ) were purchased from Kanto Chemical Co. Inc., Japan. Polyvinyl alcohol (PVA, 98.0–99.0%, hydrolyzed) was purchased from Thermo Scientific., Japan.

### 2.2. Preparation of SA/PVA hydrogel beads

The polymer solutions were prepared by dissolving 1 g of SA and 5 g of PVA in 100 mL of deionized water. PVA was initially solubilized in boiling water (approximately 100 °C) with continuous stirring until

complete dissolution was achieved. Once the PVA was completely dissolved, the solution was allowed to cool slightly before the SA was added and mixed thoroughly to achieve a homogeneous solution. After dissolution, the reaction mixture was cooled to ambient temperature. The 1 wt% SA solution was mixed with the 5 wt% PVA solution in a flask to form hydrogel beads. The mixture was shaken for 30 minutes at room temperature (25°–30°C) using a water bath shaker to integrate and create the hydrogel precursor properly. The solution was then injected dropwise into a 10 wt%  $\text{CaCl}_2$  aqueous solution previously prepared using a 10 mL syringe. The hydrogel beads were allowed to grow for 90 minutes to produce the final form before being rinsed with ethanol and distilled water and dried in an oven at 60°C for 24 hours (Fig. 1).

### 2.3. Plasma treatment

The samples were treated by a vacuum plasma machine (Sakigake, YHS-R) mounted with a vacuum pump (Ulvac Kiko inc, G-50DA) (Fig. 2a). The treatments were categorized as follows: NP (non-plasma) denotes untreated hydrogel beads, serving as the control; while P5, P10, and P15 were exposed to plasma treatment for 5 min, 10 min, and 15 min, respectively, under a vacuum chamber maintained at 0.1 bar with consistent power of 15 watts (Fig. 2b). After every treatment, the hydrogel beads were removed from the chamber and kept in airtight containers for further analysis.

### 2.4. Batch adsorption experiments

Batch adsorption experiments were conducted using a rotor (VMRC-5, As One) at an agitation speed of 100 rpm and a temperature of 30 °C. The experiments were conducted at three replicates to ensure the data's accuracy. 50 mL of a 10 mg L<sup>-1</sup> dye solution and 0.01 g of adsorbent were placed into 100 mL glass flasks.

Isotherm analyses were conducted using a constant dosage of adsorbent 0.1 g in 200 mL of aqueous solution. Initial concentrations, such as 100, 200, 300, 400, 500, and 600 mg L<sup>-1</sup>, were used at this stage [39]. While adsorption kinetics were performed using various contact durations of 5, 15, 30, 60, and 120 minutes. Adsorption thermodynamics were investigated at temperatures of 30, 40, 50, and 60 °C to assess the influence of temperature on the adsorption capacity. The MB dye removal percentage and adsorption capacity between the amounts adhere to the subsequent equations [40]:

$$\text{Dye removal (\%)} = \frac{(C_0 - C_e)}{C_0} \times 100 \quad (1)$$

$$q_e = \frac{(C_0 - C_e)}{m} V \times 100 \quad (2)$$

Where  $q_e$  signifies the mass of the dye adsorbed by the adsorbent in 1 g of adsorbent material (mg g<sup>-1</sup>),  $C_e$  is the level of concentration at equilibrium (mg L<sup>-1</sup>), whereas  $C_0$  is the initial dye concentration,  $V$  is the total volume of solution composite (L), and  $m$  is the entire mass of the adsorbent (g).

### 2.5. Determination of carboxyl groups (-COOH)

The determination of carboxyl groups on hydrogel beads is accomplished through the poem's potentiometric titration method, which has been modified slightly [41]. Dry hydrogel beads were put into 2.5 mL NaCl solution (20 mM) and left for 30 minutes. After that, HCL was added until the pH of the solution reached 3.0. Then, the solution was added with NaOH immediately until the pH reached 11.0. -COOH can be analysed quantitatively by following the following method [42]:

$$-\text{COOH}(\text{mmol/g} - 1) = \frac{V\text{NaOH} \times M\text{NaOH}}{Wd} \quad (3)$$

Where, -COOH (mmol g<sup>-1</sup>) means carboxyl content of the sample,

$V_{NaOH}$ , the volume of NaOH added to the solution until it reaches a pH of 11.0,  $M_{NaOH}$  means NaOH molarity used, and  $W_d$  refers to the initial dry weight of the hydrogel beads used.

## 2.6. Swelling capacity

The study of swelling properties was conducted under room temperature for 24 hours on a batch shaker. The swelling capacity was investigated under different pH solutions of 4, 6, 8, and 10 to reflect different ionic conditions. These swelling properties could then be ascertained from the following equation [43]:

$$\text{Swelling(\%)} = \frac{W_1 - W_0}{W_0} \times 100 \quad (4)$$

Where  $W_1$  and  $W_0$  are the weight of swollen dried hydrogel beads and the dry weight of hydrogel beads, respectively.

## 2.7. Adsorption kinetics

Adsorption kinetics are crucial for evaluating the adsorption rate, as they directly affect the overall efficiency of the adsorption process. In the present study, we used pseudo-first-order and pseudo-second-order models. 0.01 g of the adsorbent was treated in a 20 mg L<sup>-1</sup> of MB solution under ambient temperature on a batch shaker. MB solution were then collected at 5, 15, 30, 60, and 120 minutes and the solution concentrations were determined by UV-V values of the spectrophotometer. The linear adsorption capacity at that given time interval ( $q_t$ ) was calculated using the Eqs. (5) and (6) [44]:

$$qt = \frac{(C_0 - C_t)V}{m} \quad (5)$$

$$\log(q_e - q_t) = \log q_e - K_1 t \quad (6)$$

$$\frac{t}{q_t} = \frac{1}{K_2} + \frac{t}{q_e} \quad (7)$$

Where,  $q_t$  is adsorption capacity at time  $t$  (mg g<sup>-1</sup>),  $C_0$  is initial dye concentration (mg L<sup>-1</sup>),  $C_t$  is dye concentration at time  $t$  (mg L<sup>-1</sup>),  $V$  is volume of the solution (L),  $m$  is mass of the adsorbent (g),  $K_1$  (min<sup>-1</sup>) is the rate constant of the pseudo-first-order model, while  $t$  (min) represents time.  $K_2$  represents the constant rate in pseudo-second-order models.

## 2.8. Adsorption isotherms

The adsorption isotherm tests involved equilibrating adsorbent with MB solutions of starting concentrations between 100 and 600 mg L<sup>-1</sup> at room temperature. The remaining concentration of MB in the solution was quantified upon reaching equilibrium. The MB quantity adsorbed onto the hydrogel beads,  $q_e$  (mg g<sup>-1</sup>), was determined in Eq. (7) [45]:

$$q_e = \frac{q_{\max} K_L C_e}{1 + K_L C_e} \quad (8)$$

Where  $q_e$  is the equilibrium adsorption capacity (mg g<sup>-1</sup>),  $q_{\max}$  is the maximum adsorption capacity (mg g<sup>-1</sup>),  $K_L$  is the Langmuir constant, which reflects the affinity between the adsorbate and adsorbent (L mg<sup>-1</sup>), and  $C_e$  is the equilibrium concentration of MB in solution (mg L<sup>-1</sup>). The values of  $K_L$  and  $q_{\max}$  were determined through nonlinear regression. The Freundlich isotherm was determined by [46]:

$$q_e = K_F C_e^{1/n} \quad (9)$$

Where,  $K_F$  (mg g<sup>-1</sup>) is the adsorption equilibrium constant, signifying the adsorption capacity, and  $1/n$  represents the adsorption intensity. The parameters  $K_F$  and  $1/n$  were determined through the nonlinear fitting of the data.

The Temkin isotherm was expressed as:

$$q_e = B \ln(K_T C_e) \quad (10)$$

where  $B = \frac{R_T}{b}$  (J mol<sup>-1</sup>) relates to the heat of adsorption,  $K_T$  (L mg<sup>-1</sup>) is the equilibrium binding constant, is the universal gas constant (8.314 J mol<sup>-1</sup> K<sup>-1</sup>), and  $T$  is the absolute temperature (K). The Temkin model assumes that the adsorption heat decreases linearly with surface coverage.

The Redlich-Peterson isotherm was determined by:

$$Q_e = \frac{K_R C_e}{1 + \beta C_e^\gamma} \quad (11)$$

Where  $K_R$  (L g<sup>-1</sup>) is the Redlich-Peterson isotherm constant,  $\beta$  (L mg<sup>-1</sup>) is a parameter related to adsorption strength, and  $\gamma$  is an exponent that ranges between 0 and 1, representing the degree of heterogeneity of the adsorption sites. The Redlich-Peterson model combines the characteristics of both Langmuir and Freundlich isotherms.

## 2.9. Adsorption thermodynamics

By determining the standard entropy ( $\Delta S^\circ$ ), standard enthalpy ( $\Delta H^\circ$ ), and standard Gibbs free energy ( $\Delta G^\circ$ ), the thermodynamic properties of the MB dye adsorption process have been clarified. The equations defining their relationships are as follows [47]:

$$Kc = \frac{q_e}{C_e} \quad (12)$$

$$\Delta G^\circ = -RT \ln Kc \quad (13)$$

$$\ln Kc = \frac{\Delta S^\circ}{R} - \frac{\Delta H^\circ}{RT} \quad (14)$$

Where  $Kc$  is the equilibrium constant and  $R$  (8.314 J mol<sup>-1</sup>K<sup>-1</sup>) is the gas constant.  $q_e$  (mg g<sup>-1</sup>) and  $C_e$  (mg L<sup>-1</sup>) are the equilibrium concentration and adsorption amount, respectively. The slope and intercept of the plot between  $\ln Kc$  and  $1/T$  (45) were used to determine the  $\Delta H^\circ$  and  $\Delta S^\circ$ .

## 2.10. MB desorption and reusability

Following the MB adsorption process, the hydrogel beads were separated by centrifugation, dried, and subjected to desorption before reusing. The desorption process was optimized by evaluating the effects of different solvents (ethanol, distilled water, and various ethanol-distilled water ratios), desorption time (0.5–2 hours), and temperature variations (303–323 K) using a stepwise approach. Once the optimal desorption conditions were determined, the adsorbent was reused for adsorption under the optimized parameters, and the number of reuse cycles was recorded. The equations defining the desorption capacity are as follows [48]:

$$\text{Desorption capacity}(q_{de}) \text{ (mg g}^{-1}\text{)} = \frac{VC_f}{m} \quad (15)$$

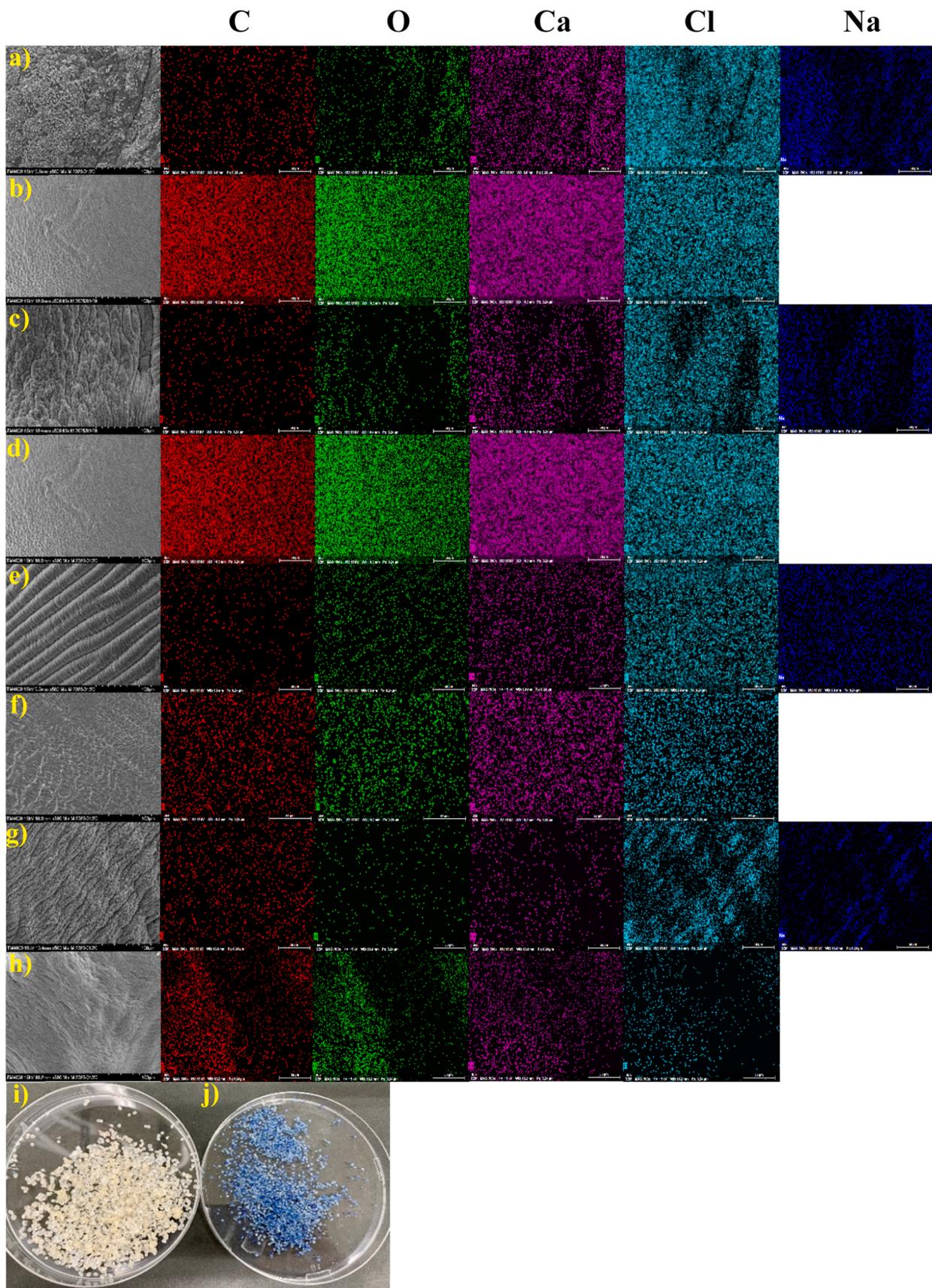
$$\text{Desorption efficiency (\%)} = \frac{C_f}{C_0} \times 100 \quad (16)$$

where  $V$  is the volume of the dye solution (L),  $m$  is the mass of the adsorbent used (mg),

$CF$  is the final concentration of MB in the solvent (mg L<sup>-1</sup>), and  $C_0$  is the initial concentration of MB (mg L<sup>-1</sup>).

## 2.11. Characterization

Surface images of the SA/PVA hydrogel beads was performed using Scanning electron microscope (SEM) linked with energy dispersive X-ray spectroscopy (EDS) Miniscope TM-4000 Plus II (Hitachi-Hitech,



**Fig. 3.** SEM EDS of hydrogel beads (a) NP (before adsorption) (b) NP (after adsorption) (c) P5 (before adsorption) (d) P5 (after adsorption) (e) P10 (before adsorption) (f) P10 (after adsorption) (g) P15 (before adsorption) (h) P15 (after adsorption). Photograph of hydrogel beads (i) before adsorption (j) after adsorption.

**Table 1**  
EDS data composition of hydrogel beads before and after adsorption.

Sample	Composition (Atom %)				
	C	O	Ca	Cl	Na
Before adsorption					
NP	40.84	21.59	5.03	18.38	14.15
P5	41.24	26.40	4.99	16.84	10.53
P10	37.77	35.29	5.09	15.10	6.75
P15	73.51	8.25	1.03	10.85	6.36
After adsorption					
NP	43.64	50.85	3.87	1.65	-
P5	46.58	48.96	3.31	1.15	-
P10	49.52	40.41	5.78	4.29	-
P15	46.84	48.70	3.63	0.83	-

Tokyo, Japan). ATR-FTIR assessed the adsorbent's functional groups from 500 to 4000  $\text{cm}^{-1}$  (Thermo Scientific Nicolet iS10, Thermo Fisher Scientific Inc., Waltham, MA, USA). The residual concentration of MB before and after the experiment was evaluated using a UV-Vis spectrophotometer (JASCO V-530) at a wavelength of 664 nm.

### 3. Result and discussion

#### 3.1. Characterization of hydrogel beads

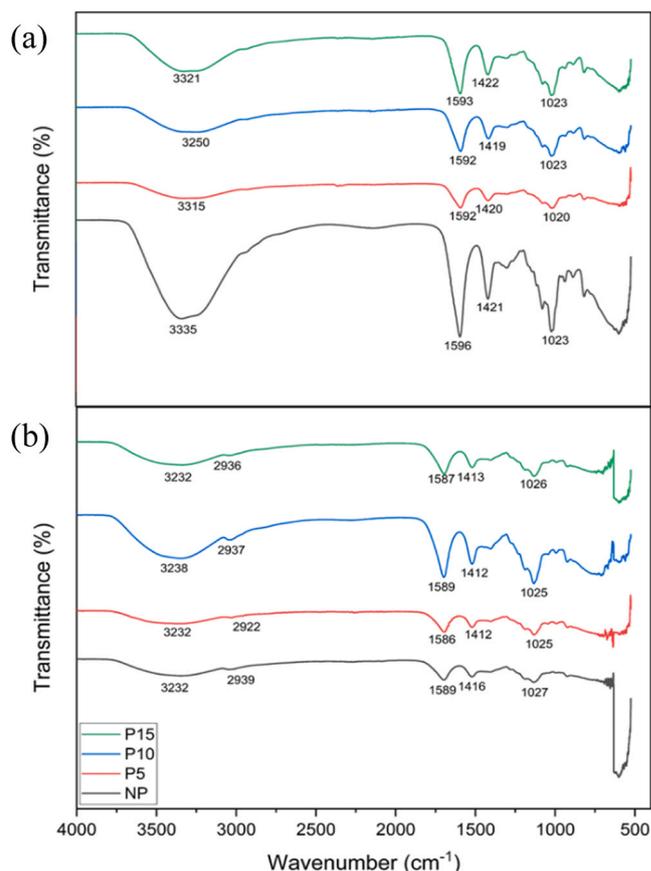
The surface morphology of the hydrogel beads before and after the adsorption process was investigated using SEM imaging (Fig. 3). The results indicated that plasma treatment caused structural changes, such as ruptures or more open porosity in the hydrogel beads (P5, P10, and P15) (Fig. 3c, e, g), compared to the untreated sample (NP) (Fig. 3a), which are believed to enhance the adsorption process (Fig. 5a). Additionally, post-adsorption images of all samples (Fig. 3b, d, f, h) showed a more pliable surface, suggesting the uptake of MB onto the hydrogel bead surfaces. These findings suggest that LPCP treatment improved the surface functionality of the hydrogel beads, facilitating efficient MB adsorption from aqueous solutions [49]. The elemental composition of hydrogel beads was analyzed using SEM-EDS to evaluate the effects of plasma treatment and adsorption processes as shown in Table 1 in before adsorption, NP treatments showed a carbon (C) content of 40.84 % and oxygen (O) content of 21.89 %, reflecting the basic composition of the hydrogel beads matrix. The presence of calcium (Ca) and chlorine (Cl) due to sample preparation used  $\text{CaCl}_2$  as crosslinking agent. Plasma-treated samples (P5, P10, P15) exhibited noticeable changes in composition, with an increase in O content (e.g., P5 and P10: 26.40 % and 35.29 %, respectively), and a reduction in Cl content. This result is consistent with a previous study [50,51], which showed that the elemental O content of the sample increased after plasma irradiation, possibly due to the introduction of O from the plasma source during the plasma crosslinking process. The prolonged plasma treatment in P15 resulted in the highest C content (73.51 %) and the lowest O content (8.25 %), suggesting enhanced surface carbonization or structural modifications caused by extended plasma exposure [52].

After adsorption, all samples exhibited significant increases in O content, indicating the incorporation of oxygenated species through adsorption [53]. For instance, the O content in NP increased from 21.59 % to 50.85 %, and in P15, it increased from 8.25 % to 48.70 %. Simultaneously, the Cl content decreased drastically across all samples (e.g., NP: 18.38–1.65 %, P5: 16.84–1.15 %, P10: 15.10–4.29 %, and P15: 10.85–0.83 %), suggesting ion exchange or replacement of Cl by other species during adsorption. Notably, C content also increased in all samples after adsorption, highlighting the deposition or interaction of organic molecules on the hydrogel surfaces [54]. P15 exhibited the most significant changes among the plasma treated samples, demonstrating its enhanced surface functionalization and adsorption capacity due to prolonged plasma exposure [55]. Moreover, the photograph of hydrogel

**Table 2**  
Swelling properties and -COOH value of the hydrogel beads.

Treatment	Swelling (%)	-COOH ( $\text{mmol g}^{-1}$ )
NP	14.14 $\pm$ 0.13 <sup>a</sup>	12.28 $\pm$ 0.02 <sup>c</sup>
P5	21.01 $\pm$ 0.07 <sup>b</sup>	6.78 $\pm$ 0.01 <sup>b</sup>
P10	24.78 $\pm$ 0.1 <sup>c</sup>	6.06 $\pm$ 0.02 <sup>b</sup>
P15	57.66 $\pm$ 0.12 <sup>d</sup>	5.08 $\pm$ 0.01 <sup>a</sup>

Mean values in the same column followed by different letters indicate significant differences ( $P < 0.05$ ) according to Duncan's test.

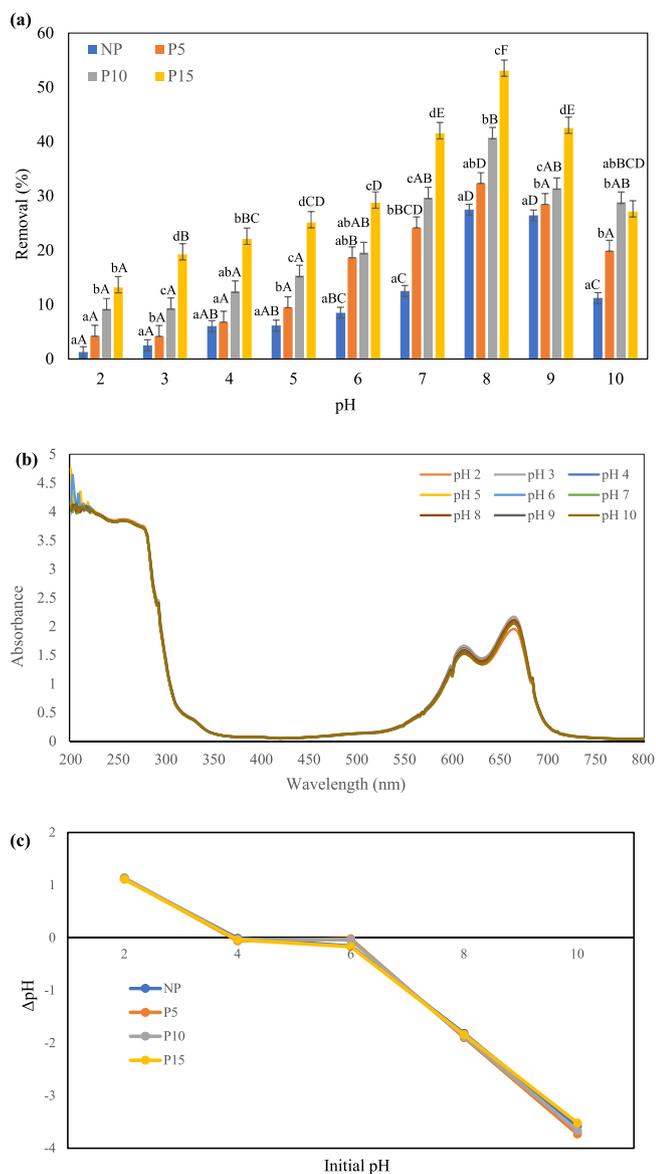


**Fig. 4.** FTIR spectra of hydrogel beads (a) before and (b) after adsorption of MB.

beads before and after adsorption as shown in Fig. 3i-j. The color was changed to blue after adsorption, indicating the successful the adsorption process.

Table 2 demonstrates the effect of LPCP treatments on the physico-chemical properties of the hydrogel beads. Statistical analyses show that LPCP treatments exhibited high significance for swelling properties ( $P < 0.05$ ), in contrast with the COOH group ( $P < 0.05$ ). It might have caused the decomposition of oxygen functional groups from the damaged hydrogel beads surface [56]. The results was similar to those reported previously [57], that used plasma treatment on fiber composites.

These results confirm the proposed mechanisms of LPCP surface modification, whereby high energy interacts and fuses with the hydrogel material, which leads to physical reconstruction (more surface area for interaction and increased surface roughness) and chemical functionalization (incorporation of polar hydroxyl, carboxyl, and carbonyl groups on the hydrogel surface) [58]. These will likely result in more enhanced hydrophilicity at the surface and greater density of active functional groups to facilitate better adsorption performance of SA/PVA hybrid hydrogel beads for water treatment, particularly the elimination of



**Fig. 5.** (a) Effect of Initial pH on MB removal percentage. (b) UV-Vis Spectrum of MB Oligomer Formation (c)  $pH_{zpc}$  values of hydrogel beads. Different uppercase letters (A–C) on the bars within the same pH indicate significant differences. Different lowercase letters (a–b) on the bars within the same treatment indicate significant differences ( $P < 0.05$ ).

cationic dyes, such as MB [59].

The FTIR spectra of the hydrogel beads as shown in Fig. 4. The data indicate changes in functional group peaks, namely in hydroxyl (O–H), carbonyl (C=O), and C–O–C bonds, which signify structural and bonding modifications caused by plasma exposure. The peak changes from  $3335\text{ cm}^{-1}$  to  $3321\text{ cm}^{-1}$  indicate O–H stretching, typically associated with hydroxyl groups [60]. Plasma treatment shifts these peaks to slightly lower wavenumbers, indicating a change in hydrogen bonding or interactions among hydroxyl groups. This transition suggests plasma treatment may modify the surface chemistry by strengthening hydrogen bonding or adding new surface hydroxyl groups [61]. The peaks around  $1596\text{--}1592\text{ cm}^{-1}$  indicate C=O stretching in carboxylate groups, and exhibit minor shifts. The changes identified in this location indicate that plasma treatment may amplify these groups reactivity, thereby affecting the material's adsorption characteristics. Peaks between  $1421$  and  $1304\text{ cm}^{-1}$ , associated with C–O stretching and O–H bending vibrations, exhibit variations, signifying alterations in the structure of carboxylate

and hydroxyl groups. Plasma treatment likely influences the orientation or bonding of these groups, so reinforcing the notion that the surface chemistry is being altered. In the C–O–C stretching zone ( $1079\text{--}1023\text{ cm}^{-1}$ ), crucial for the PVA and alginate backbone, variations in peak intensity and minor shifts in wavenumber are seen. This suggests that the plasma treatment causes rearrangements or cross-linking within the polymer foundation, altering the hydrogels structural characteristics [62]. The lower wavenumber range (below  $1000\text{ cm}^{-1}$ ) encompasses peaks associated with C–H bending and deformation modes, with intensity variations presumably indicating modifications in the hydrogel matrix due to plasma treatment [63]. The changes in peak positions and intensities indicate that plasma treatment continuously changes the hydrogel surfaces functional groups, with the modification amount increasing with longer exposure times. These changes improve the hydrophilicity and reactive sites on the hydrogel surface, potentially improving its adsorption capacity [64].

After adsorption, significant changes are observed in the plasma-treated and non-treated samples (Fig. 4b). The –OH stretching peaks shift to lower wavenumbers (e.g.,  $3232\text{--}2937\text{ cm}^{-1}$ ) and decrease in intensity, indicating the involvement of hydroxyl groups in MB adsorption. The carboxylate peaks ( $1596\text{--}1587\text{ cm}^{-1}$  and  $1421\text{--}1412\text{ cm}^{-1}$ ) also shift and decrease, suggesting an interaction between the  $\text{COO}^-$  groups in the SA and the cationic MB dye. Furthermore, the peaks around  $1025\text{--}1027\text{ cm}^{-1}$  show changes in transmittance, confirming that the adsorption process involves the C–O and C–O–C functional groups. Plasma-treated samples, particularly P15, display more pronounced spectral changes than NP, indicating enhanced interaction with MB due to the plasma-induced functional group activation. These results suggest that cold plasma treatment improves the hydrogel surface properties, enhancing its ability to adsorb MB efficiently [65].

### 3.2. Initial pH effect

The pH-dependent adsorption behavior of untreated and plasma-treated adsorbents was systematically investigated over a pH range of 2–10. The results (Fig. 5a) indicate that the adsorption efficiency increased with pH, reaching a maximum of 8.2 for all samples. The highest removal efficiency was observed for P15 (53.03%), followed by P10 (40.72%), P5 (32.29%), and NP (27.48%). This trend suggests that plasma treatment enhances the adsorption capacity of hydrogel beads by modifying surface functional groups, thereby increasing interaction sites for MB. At lower pH values (2.0–4.0), adsorption was significantly reduced, with removal efficiencies below 10% for NP and P5, while P10 and P15 showed slightly higher efficiencies (up to 22.08% for P15 at pH 4). The decrease in adsorption efficiency at acidic conditions is attributed to electrostatic repulsion between the cationic MB species and the protonated functional groups of the hydrogel beads [66].

The  $pK_a$  of MB is approximately 3.8 [67], which influences its molecular speciation and adsorption behavior. At pH values below this range, MB predominantly exists in its monomeric cationic form ( $\text{MB}^+$ ), leading to strong electrostatic repulsion from the positively charged hydrogel surface. Consequently, limited adsorption occurs due to reduced attractive interactions [68]. As pH increases beyond the  $pK_a$ , MB deprotonates slightly, reducing electrostatic repulsion and facilitating stronger interaction with negatively charged hydrogel surfaces. This is particularly evident in the significant increase in removal efficiency at pH 6.0–8.0, where enhanced electrostatic attraction between MB and surface functional groups (e.g., carboxyl and hydroxyl groups) plays a dominant role [69]. The optimal adsorption observed at pH 8.2 suggests that MB is predominantly in its monomeric form at this pH, maximizing adsorption through electrostatic and hydrogen bonding interactions. At alkaline pH conditions ( $\geq 9.0$ ), the adsorption efficiency decreased, likely due to the formation of MB oligomers. UV-Vis spectral analysis in the range of 200–800 nm confirmed this phenomenon

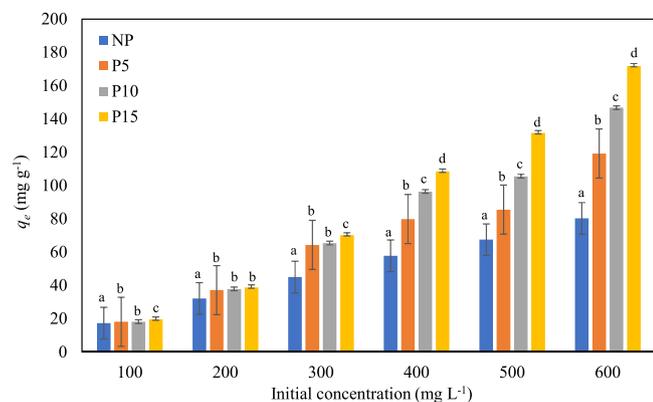


Fig. 6. Effect of initial concentration on MB removal percentage. Different lower-case letters indicate a significant difference in means ( $P < 0.05$ ).

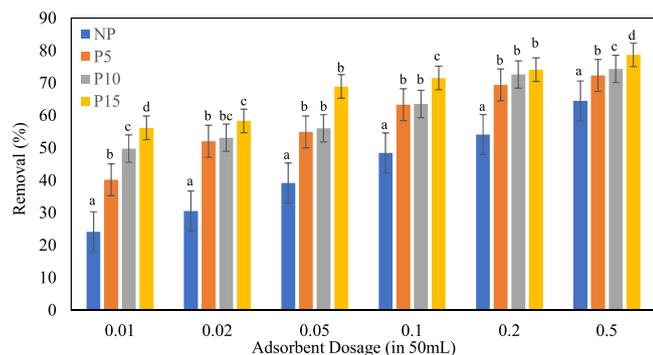


Fig. 7. Adsorbent dosage effects on the removal percentage of MB. Different lower-case letters indicate a significant difference in means ( $P < 0.05$ ).

(Fig. 5b), as additional absorbance peaks at  $\sim 570$  nm and 620–660 nm correspond to dimeric and aggregated MB species [70]. The formation of MB oligomers affects adsorption efficiency in two main ways: steric hindrance from larger molecular aggregates reduces their accessibility to adsorption sites, limiting interaction with the hydrogel surface. Second, the aggregation alters charge distribution, reducing electrostatic attraction between MB and the hydrogel, thereby decreasing adsorption efficiency at high pH values [71].

pH<sub>ZPC</sub> analysis (Fig. 5c) further supports these observations, where a positive pH shift at low initial pH values (2.0–4.0) suggests significant uptake of H<sup>+</sup> ions by the hydrogel surface, whereas a negative pH shift at pH 8.0–10.0 indicates the release of H<sup>+</sup> ions into the solution. This amphoteric behavior is attributed to the presence of surface functional groups such as carboxyl (-COOH) and hydroxyl (-OH), which undergo protonation and deprotonation depending on the surrounding pH [72]. The consistent pH shift trends across plasma-treated and untreated samples suggest that plasma treatment enhances surface functionalization without significantly altering the isoelectric point of the hydrogel beads.

### 3.3. Impact of the initial MB concentration

Fig. 6 illustrates the effect of the initial methylene blue (MB) concentration on the adsorption capacity for different plasma treatments (NP, P5, P10, and P15). The adsorption capacity increased with increasing initial MB concentration, indicating that higher concentrations provide a stronger driving force for overcoming mass transfer resistance between the solution and the adsorbent surface. This trend reflects the enhanced availability of adsorption sites and improved surface interactions following plasma treatment. The plasma treated beads (P5, P10, and P15) consistently exhibited higher adsorption

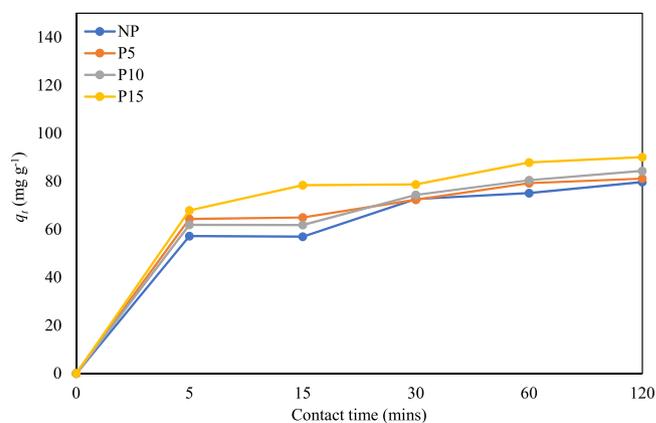


Fig. 8. Effect of contact time on removal percentage of MB.

capacity than the untreated beads (NP), confirming that plasma modification enhances surface functionality and active site availability. At the highest MB concentration ( $600 \text{ mg L}^{-1}$ ), the adsorption capacity reached  $175.506 \text{ mg g}^{-1}$  for P15, which was significantly higher than the untreated beads (NP) at  $108.568 \text{ mg g}^{-1}$ . Statistical analysis using different lowercase letters indicates significant differences in adsorption capacity among the treatments ( $P < 0.05$ ). The increase in adsorption capacity with plasma treatment can be attributed to enhanced surface roughness, increased pore volume, and improved surface chemistry, which provide more active sites for MB adsorption [73]. The improved adsorption performance of the plasma-treated beads highlights plasma modifications effectiveness in enhancing the hydrogel beads adsorption capacity and surface interactions.

### 3.4. Effect of adsorbent dosage

Fig. 7 shows the effect of adsorbent dosage on MB removal efficiency. The results indicate that increasing the adsorbent dosage enhances removal efficiency, consistent with typical adsorption system behavior, where higher adsorbent concentrations provide more active sites, improving adsorption effectiveness [74]. At the lowest adsorbent dosage (0.01 mg in 50 mL solution), the P5 treatment exhibited the highest removal efficiency ( $\sim 50\%$ ), suggesting that plasma treatment enhances the surface characteristics of the beads, likely by increasing the surface area or introducing functional groups that promote interaction with MB molecules [75]. At the highest adsorbent dosage (0.5 mg), the P15 sample achieved the highest removal percentage ( $\sim 80\%$ ), indicating that increased surface modification leads to greater adsorption site availability until the beads reach their saturation point [76]. LPCP-treated samples (P5, P10, P15) exhibited higher removal efficiency across all adsorbent dosages compared to untreated samples (NP), confirming the positive effect of plasma treatment on adsorption performance.

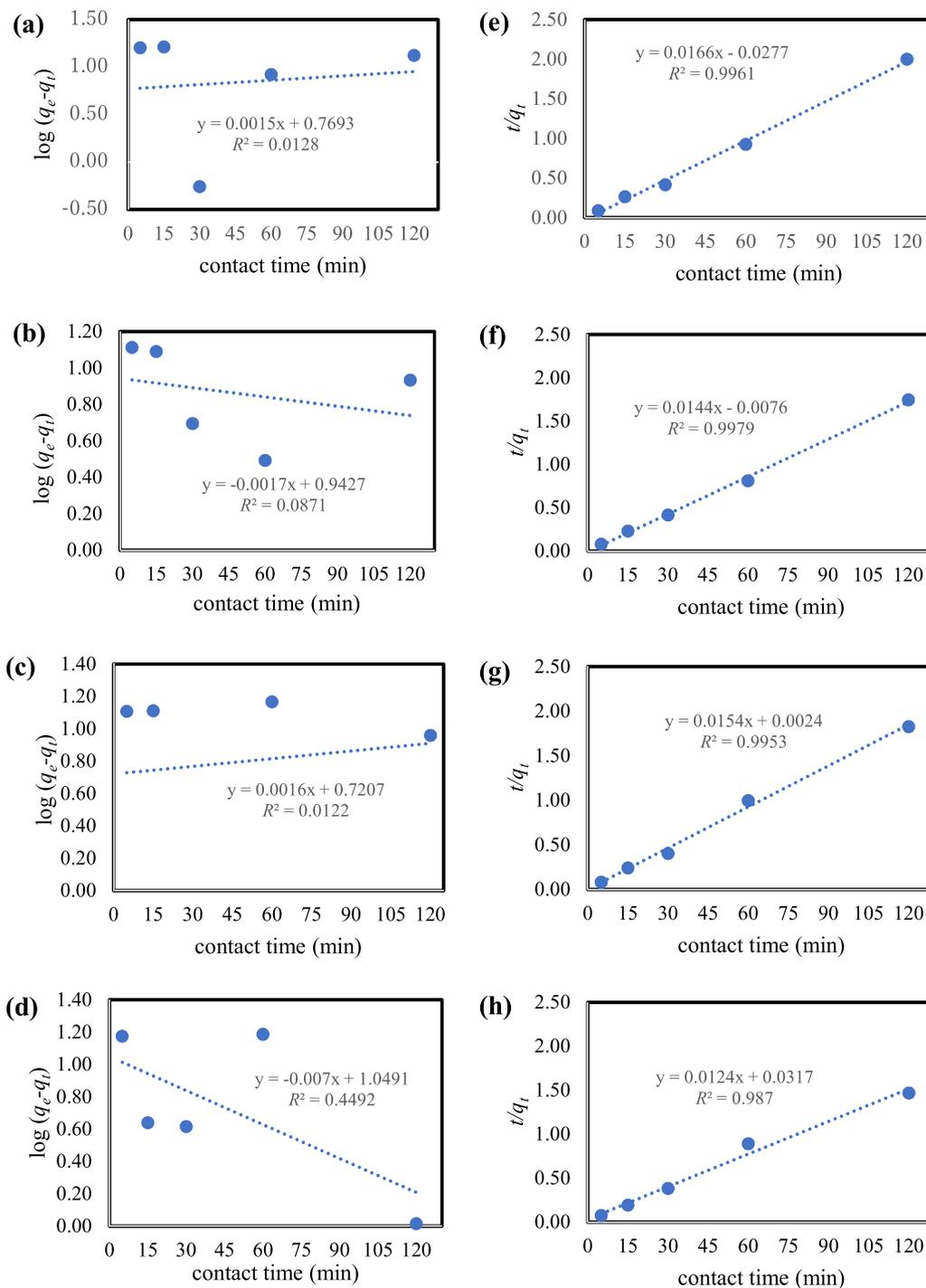
Fig. 8 shows the effect of contact time on the removal efficiency of MB. The removal efficiency generally increases with contact time from 5 to 120 minutes, with the highest values observed at 120 minutes. This trend is consistent with typical adsorption behavior, where prolonged contact time enhances the interaction between the adsorbate (MB) and adsorbent (hydrogel beads), leading to greater removal efficiency [77].

The initial increase is most pronounced between 5 and 30 minutes, likely due to the high availability of adsorption sites at the beginning of the process. After 30 minutes, the increase becomes more gradual as the adsorption sites begin to reach saturation [78]. Plasma-treated beads (P5, P10, P15) exhibited higher removal efficiencies than untreated beads (NP) at all contact times, with the highest removal observed for the P15 sample, indicating that enhanced surface modification increases the adsorption capacity.

The adsorption kinetics of MB onto hydrogel beads were analyzed

**Table 3**  
Kinetic parameters for the adsorption of MB onto hydrogel beads.

Kinetic models	Parameters	Treatment			
		NP	P5	P10	P15
Pseudo-first-order	$qe$ ( $mg\ g^{-1}$ )	2.158	2.567	2.056	2.855
	$K_1$ ( $g\ mg^{-1}\ min^{-1}$ )	0.0000125	0.0000142	0.0000133	0.0000583
	$R^2$	0.013	0.087	0.012	0.449
Pseudo-second-order	$qe$ ( $mg\ g^{-1}$ )	60.241	69.444	64.935	80.645
	$K_2$ ( $g\ mg^{-1}\ min^{-1}$ )	0.010	0.027	0.099	0.005
	$R^2$	0.996	0.998	0.995	0.987



**Fig. 9.** Adsorption kinetics using (a)NP, (b) P5, (c) P10, (d) P15 pseudo-first-order and (e) NP, (f) P5, (g) P10, and (h) pseudo-second-order curves.

**Table 4**  
Isotherm parameters for the adsorption of MB onto hydrogel beads.

Isotherm models	Parameters	Treatment			
		NP	P5	P10	P15
Langmuir	$q_{max}$ (mg g <sup>-1</sup> )	108.568	112.315	147.900	175.506
	$K_L$ (L mg <sup>-1</sup> )	1.224	2.178	2.440	5.364
	$R_L$	0.0081	0.0046	0.0041	0.0019
	$R^2$	0.954	0.932	0.903	0.858
Freundlich	$K_F$ (g mg <sup>-1</sup> min <sup>-1</sup> )	27.673	143.193	141.595	530.264
	$1/n$	0.552	0.453	0.484	0.456
	$R^2$	0.998	0.933	0.904	0.961
	$R^2$	0.947	0.904	0.878	0.844
Temkin	$B$	22.174	20.919	29.251	30.120
	$K_T$	0.133	0.305	0.231	0.674
	$R^2$	0.947	0.904	0.878	0.844
Redlich-peterson	$\beta$	0.0088	0.0229	0.0126	0.01298
	$K_R$	1.057	2.49	2.028	3.009
	$R^2$	0.977	0.839	0.842	0.893

**Table 5**  
Thermodynamic adsorption of MB onto Hydrogel Beads.

	$\Delta G^\circ$ (kJ mol <sup>-1</sup> )				$\Delta H^\circ$ (kJ mol <sup>-1</sup> )	$\Delta S^\circ$ (kJ mol <sup>-1</sup> K <sup>-1</sup> )
	303	313	323	333		
NP	-2.462	-2.643	-3.107	-7.553	37.210	128.767
P5	-2.660	-4.025	-4.387	-6.629	27.923	101.165
P10	-2.680	-2.913	-3.567	-5.923	22.751	82.902
P15	-3.782	-4.842	-6.155	-5.131	7.986	40.336

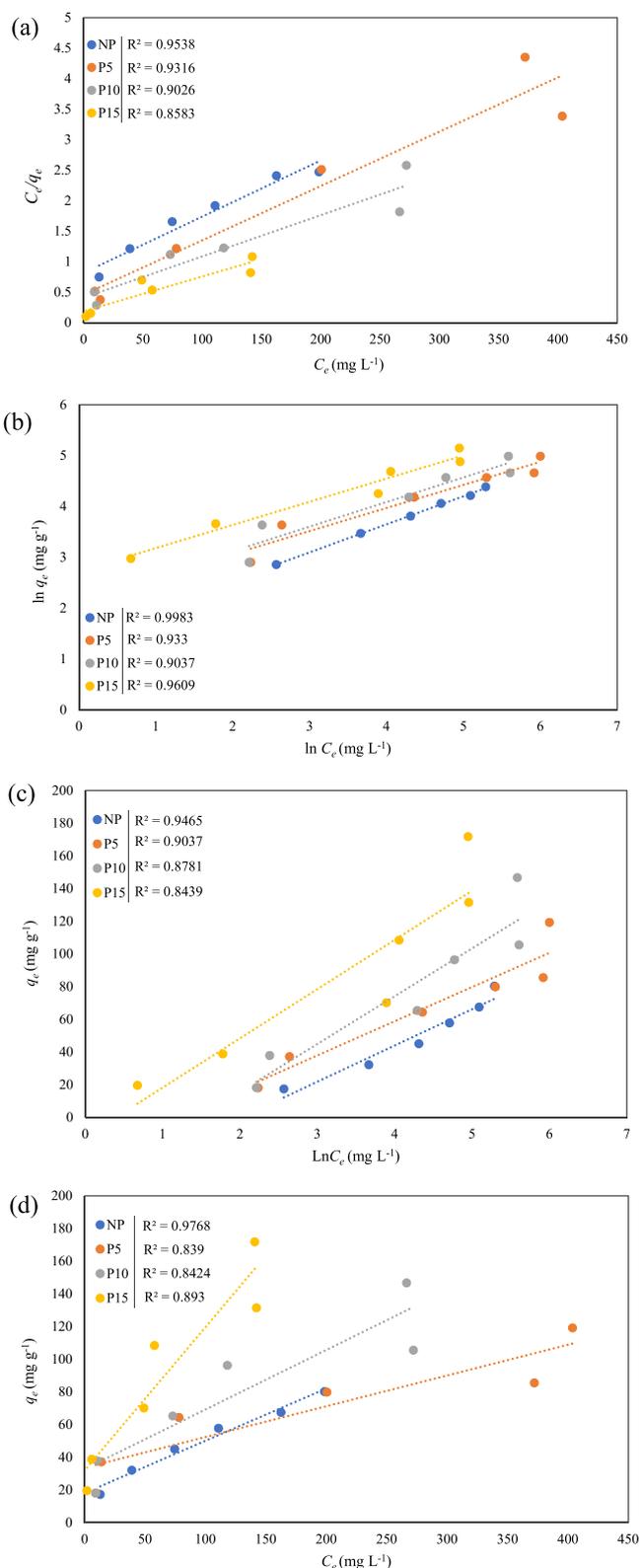
using pseudo-first-order and pseudo-second-order models. Table 3 demonstrates that the pseudo-second-order model provided a more accurate representation of the experimental data for all samples, shown by the higher  $R^2$  values ranging from 0.987 to 0.998, compared to the pseudo-first-order model. This suggests chemisorption is probably the rate-limiting phase [79], as illustrated in Fig. 9.

The equilibrium adsorption capacity increased with the duration of plasma treatment, attaining its highest point of 80.645 mg g<sup>-1</sup> at P15, with the lowest for the untreated sample (NP) at 60.241 mg g<sup>-1</sup>. The improvement in adsorption capacity can be ascribed to enhanced surface characteristics, presumably resulting from improved functional group availability and modifications in surface area due to plasma exposure [80].

### 3.5. Adsorption isotherm analysis

Table 4 presents the linear regression coefficients ( $R^2$ ) alongside key adsorption isotherm parameters for each treatment, highlighting the performance of the Langmuir and Freundlich models. The Freundlich model consistently exhibits superior  $R^2$  values across all treatments (NP, P5, P10, and P15), with values ranging from 0.904 to 0.998, indicating a better fit for the experimental adsorption data. This suggests that the adsorption process may partially occur in a multilayer configuration under these conditions. The extended plasma treatment likely induces surface irregularities and heterogeneities, enhancing the complexity of the adsorption surface [81]. This is reflected in the significant increase in the Freundlich constant  $K_F$ , which rises from 27.673 (NP) to 530.264 (P15), indicating higher adsorption capacity with increased plasma treatment. Consequently, the Freundlich model, which accounts for non-uniform adsorption sites and multilayer adsorption, provides a more accurate description of the adsorption process [82]. (Table 5)

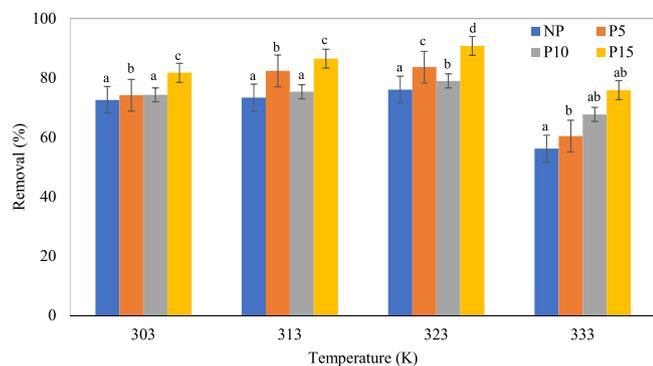
The Langmuir model shows a progressive increase in maximum adsorption capacity ( $q_{max}$ ) from 108.568 mg g<sup>-1</sup> (NP) to 175.506 mg g<sup>-1</sup> (P15), reflecting enhanced monolayer adsorption capacity with plasma treatment. The increase in  $q_{max}$  suggests that plasma treatment enhances the availability of active adsorption sites, possibly



**Fig. 10.** Adsorption isotherm models of Langmuir (a), Freundlich (b), Temkin (c), and Redlich-peterson (d) for MB adsorption onto hydrogel beads.

through surface activation or modification. However, the lower Langmuir regression coefficients ( $R^2$ ) compared to the Freundlich model suggest that a heterogeneous surface model better represents the adsorption process rather than a uniform monolayer (Fig. 10).

Furthermore, the Temkin and Redlich-Peterson models were also



**Fig. 11.** Removal percentage of MB on different temperature. Different lower-case letters indicate a significant difference in means ( $P < 0.05$ ).

evaluated to gain deeper insights into the adsorption mechanism [83]. The Temkin model considers the effect of indirect adsorbate–adsorbent interactions, while the Redlich–Peterson model combines features of the Langmuir and Freundlich models, capturing monolayer and multilayer adsorption characteristics. The Temkin constant and Redlich–Peterson constant

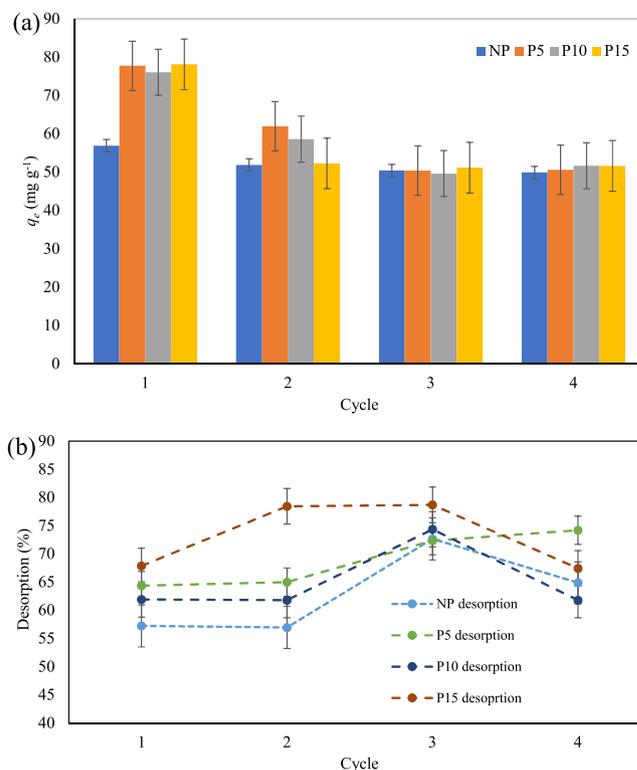
$K_R$  increased with plasma treatment, supporting the hypothesis that surface heterogeneity and the number of available binding sites increased after plasma exposure. These findings align with the adsorption dynamics physical nature, emphasizing the adsorbent surfaces heterogeneous and multi-layered characteristics, particularly following plasma treatment [84]. Enhanced surface roughness and active site availability following plasma treatment likely contribute to the improved adsorption capacity and model fit observed in the Freundlich and Redlich–Peterson models.

### 3.6. Adsorption thermodynamics

Adsorption efficiency rises with temperature from 303 K to 323 K for most treatments, indicating improved adsorbate–adsorbent interaction at high temperatures (Fig. 11). A reduction in removal effectiveness is observed at 333 K, which may be attributed to the desorption effect or a decreased affinity of adsorbates at higher temperatures. For instance, the Gibbs free energy for the NP treatment decreases from  $-2.462 \text{ kJ mol}^{-1}$  at 303 K to  $-7.553 \text{ kJ mol}^{-1}$  at 333 K. This highlights the pronounced impact of elevated temperatures on adsorption efficiency, potentially due to weakened interactions between the adsorbent and adsorbate at higher thermal conditions [85].

This discovery aligns with the endothermic characteristics of adsorption, wherein increasing temperatures initially enhance diffusion and adsorption but may ultimately reduce surface adsorption forces [86, 87]. The cold plasma treatment enhances the materials adsorption characteristics with extended exposure times, such as P15, providing maximum removal effectiveness at all temperatures. This improvement may arise from augmented surface functionality such as carboxyl and hydroxyl groups or enhanced porosity achieved through plasma treatment, as indicated in prior research on adsorbent modifications [88].

The thermodynamic analysis shows that plasma treatment improves the adsorption characteristics of hydrogel beads. The negative Gibbs free energy ( $\Delta G^\circ$ ) values for all samples signify spontaneous adsorption, with plasma-treated beads, particularly P15, exhibiting enhanced spontaneity, indicating that plasma treatment augments adsorption favorability (Table 4). The positive enthalpy ( $\Delta H^\circ$ ) values for NP, P5, and P10 indicate an endothermic process wherein adsorption is facilitated by heat absorption. Nevertheless,  $\Delta H^\circ$  values diminish with increased LPCP treatment, indicating enhanced adsorption with less heat exchange, presumably resulting from plasma-induced structural alterations in the hydrogel [89,90]. Structurally, LPCP treatment likely increases surface smoothness and reduces irregularities, making the adsorption sites more



**Fig. 12.** (a) Reusability and (b) desorption efficiency of the hydrogel beads for MB removal from water.

**Table 6**

Comparison adsorbents for MB adsorption.

Adsorbent	$q$ ( $\text{mg g}^{-1}$ )	Reference
NP	108.57	This work
P5	112.32	This work
P10	147.9	This work
P15	175.51	This work
Au-NP loaded on activated carbon	40.65	[93]
$\text{Co}_3\text{O}_4/\text{SiO}_2$ nanocomposite	53.87	[46]
PAN-g-Alginate	3.51	[94]
H. cannabinus-g-PAA/PAAM	7	[95]
Carboxylate alginate	14	[96]
PVA/Bent hydrogel	27.90	[97]
Chitosan/organic rectorite composite	10.91	[98]
CS50SP50 composite adsorbent	40.98	[97]
Raw Ball clay	34.65	[99]

uniform. It enhances surface polarity and functionality, facilitating stronger and more specific interactions with adsorbates. This improved organization reduces the randomness (entropy) of the adsorption process, as adsorbates bind more systematically to these optimized sites. LPCP treatment enhances the specific surface area and pore distribution of SA/PVA while increasing the density of functional groups, such as oxygen-containing groups. These structural changes promote a more predictable and ordered adsorption process, as evidenced by the reduction in  $\Delta S^\circ$  [91].

### 3.7. Reusability of SA/PVA hydrogel beads

The reusability data is shown in Fig. 12. Generally, plasma-treated hydrogel beads reveal high adsorption efficiency compared to the untreated sample. It might be indicated that even up to three cycles of the adsorption process, the hydrogel beads-treated plasma showed a good performance for MB removal from water. Furthermore, desorption data also showed a good performance for hydrogel beads treated plasma

compared to untreated sample. This is due to heightened surface roughness, more excellent hydrophilicity, and the incorporation of supplementary functional groups on the hydrogel surface. These alterations enhance the availability of active sites for dye molecule binding and improve dye removal efficiency [92].

### 3.8. Comparison of the maximum capacities with the previous adsorbents in literature

We compared our findings with previous research on the adsorption of MB from aqueous solutions using hydrogel beads, as summarized in Table 6. According to the literature, the P15 sample demonstrated superior performance. This suggests that the plasma exposure time plays a crucial role in enhancing the efficiency of MB removal from water.

## 4. Conclusion

Sodium alginate/polyvinyl alcohol (SA/PVA) hydrogel beads modified with low-pressure cold plasma (LPCP) were effectively utilized for the adsorption of methylene blue (MB) from aqueous solutions. The LPCP treatment significantly enhanced the beads' adsorption capacity, with the maximum adsorption capacity increasing from 108.568 mg g<sup>-1</sup> (NP) to 175.506 mg g<sup>-1</sup> (P15). Freundlich isotherm analysis showed that the adsorption process followed a multilayer mechanism, with the Freundlich constant ( $K_F$ ) increasing from 27.673 (NP) to 530.264 (P15). Kinetic analysis confirmed that the adsorption followed a pseudo-second-order model ( $R^2 > 0.98$ ), indicating chemisorption as the dominant mechanism. Thermodynamic analysis demonstrated that the process was spontaneous ( $\Delta G^\circ = -7.55 \text{ kJ mol}^{-1}$  at 333 K) and endothermic, with increased adsorption at higher temperatures. After three cycles, plasma-treated hydrogel beads maintained over 85 % adsorption efficiency, highlighting their reusability and stability. These results underscore the potential of LPCP-treated hydrogel beads as an efficient and sustainable approach for wastewater treatment, warranting further research for broader environmental applications.

### CRedit authorship contribution statement

**Hidayat Endar:** Writing – review & editing, Methodology, Formal analysis, Data curation, Conceptualization. **Wijaya Rizza:** Writing – original draft, Methodology, Investigation, Formal analysis, Data curation, Conceptualization. **Mitoma Yoshiharu:** Supervision. **Harada Hiroyuki:** Supervision. **Samitsu Sadaki:** Supervision, Data curation. **Yonemura Seiichiro:** Writing – review & editing, Supervision, Investigation, Formal analysis, Data curation, Conceptualization.

### Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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### Data availability

Data will be made available on request.

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